Dynamics of Nitrifiers in Soils of Intensively Vegetable Cultivated Areas in Sri Lanka

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ABSTRACT: Nitrification tend to reduce the agronomic fertilizer use efficiency in cropping systems because nitrate, the end product of nitrification, has high potential to loss from soil environment due to denitrification and leaching. The population characteristics of ammonia oxidizing and nitrite oxidizing microorganisms carrying out nitrification in soils may affect the effectiveness of approaches taken to suppress nitrification. A study was conducted to assess the activity and abundance of nitrifiers in soils of intensively vegetable grown regions in Sri Lanka. Soil samples were collected from 72 locations across Nuwara Eliya, Marassana, Kalpitiya and Gannoruwa representing vegetable cultivated fields managed conventionally (n=45) and organically (n=9), and uncultivated areas (n=18). Basic soil characteristics were determined using standard procedures. Potential nitrification rate (PNR) and the abundance of ammonia oxidizers and nitrite oxidizers were estimated. Chemical characteristics of soils varied widely with pH, EC and organic C% ranging from 3.8 to 8.5, 0.04 to 0.94 dS/m and 0.9% to 4.5%, respectively. The PNR of the studied soils ranged from 0.18 to 15.80 NO₃-N/kg/h. The abundance of ammonia oxidizers and nitrite oxidizers ranged from 1.96 to 5.97 log₁₀CFU g⁻¹ and 1.36 to 5.63 log₁₀CFU g⁻¹, respectively. The potential activity of nitrifiers did not correlate with the abundance of ammonia oxidizers or nitrite oxidizers. Thus, the functional and compositional diversities of nitrifying communities may be different across the soils. PNR values for studied soils are higher compared to reported values in literature. Hence, appropriate measures need to be taken to suppress nitrification as high nitrification rates could lead to reduce fertilizer use efficiency and increase risk of groundwater contamination with nitrate.

Keywords: Fertilizer use efficiency, potential nitrification rate, soil nitrifiers, vegetable cultivation

INTRODUCTION

Nearly 11 % of the N-fertilizers (urea and ammonium sulphate) imported to Sri Lanka is being consumed by vegetable sector, only second to tea sector (27 %) from among the systems that dominates cultivation under unsaturated soil moisture conditions (personal communication with National Fertilizer Secretariat, 2017). Nitrification, conversion of NH₄⁺ to NO₃⁻, is a biological transformation occurring predominantly under unsaturated soil moisture conditions since it is mediated by aerobic microorganisms. Thus, fertilize-N is converted to NO₃⁻-N in soils at different rates depending upon the dynamics of soil nitrifiers. Krishnapillai (1979)

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reported that about 35 - 50% of the ammonium and urea fertilizer applied to soils in tea plantations in Sri Lanka were nitrified within 3 to 4 weeks after application and the produced NO₃- rapidly leached from soil under high rainfall condition. In Sri Lanka, Nuwara Eliya, Kalpitiya and Marassana are some of the intensively vegetable grown regions that experience heavy application of nitrogen fertilizers. According to Jayasingha *et al.* (2011) NO₃-N in 50% of the studied groundwater samples collected from Kalpitiya exceeded WHO standards for drinking water (10 mg L⁻¹) both in dry season and rainy season with values ranging from 0.20–212.4 mg L⁻¹. A number of studies suggest association of NO₃-N in groundwater with fertilizer-N usage for vegetable cropping in the region threatening human health (Henegama *et al.*, 2013; Kuruppuarachchi, 2010; Jayasingha *et al.*, 2011; Liyanage *et al.*, 2000). However, little or no information is available on the diversity and dynamics of nitrifiers inhabiting soils of Sri Lanka.

Biological nitrification was first explained by Winogradsky during 1889-1890 describing the role of nitrifying bacteria. Two chemoautotrophic groups of bacteria broadly known as ammonia oxidizers (e.g. Nitrosomonas spp., Nitrosospira spp., Nitrosococcus spp., etc.) and nitrite oxidizers (Nitrobacter spp., Nitrospira spp., Nitrococcus spp., etc.) are known to oxidize ammonia into nitrite and subsequently nitrite into nitrate (Schmidt, 1982). However, presence of ammonia oxidizing archaea (AOA) in soil have been detected and several studies suggest that AOA may play an important role in nitrification under low ammonium concentrations and/or under suboptimal conditions defined for ammonia oxidizing bacteria (AOB), such as acidic pH, high EC, and low temperature (Leininger et al., 2006; Bernhard et al., 2010; Hofferle et al., 2010; Posser and Nicol, 2012; Hu and He, 2014). Further complete oxidation of ammonia to nitrate by a single organism via a process known as complete ammonia oxidation or comammox has been detected in some members of Nitrospira genus although the responsible microorganisms have not been isolated (Kessel et al., 2015). Organisms exhibiting comammox are considered to be having slow growth rates. Beeckman et al. (2018) also emphasized about ammonia oxidizing Achaea and comammaox bacteria who play a critical role in nitrification. Decaying organic matter, biologically fixed NH₄⁺ and added fertilizers provide NH₄+-N, the substrate for nitrification (Sylivia et al., 2005). Ammonia oxidizing bacteria seems to govern nitrification under NH₄⁺-N rich soil conditions such as in soils that received fertilizer-N, polluted with sewage sludge, or experienced leakages from septic tanks (Leininger et al., 2006; Höfferle, et al., 2010).

Both NO₃⁻ and NH₄⁺ are considered as plant available forms of N though NO₃⁻ is a mobile ion subjected to leaching and loss from soil through denitrification. Hence, nitrification rate determines how quick the N is becoming unavailable for crop uptake when provides NH₄⁺ through fertilizers in agriculture field. Potential nitrification rate (PNR) is a measure used to identify rapidity of nitrification with the provision of ample amount of NH₄⁺ which converts NH₄⁺ into NO₃⁻ in the presence of nitrifying communities for a particular soil. Enumeration of ammonia oxidizers and nitrite oxidizers coupled with most probable number (MPN) technique is commonly used to estimate the abundance of these two groups (Prosser and Nicol, 2012; Hesselsoe *et al.*, 2001). Both PNR and MPN techniques do not discriminate between archaeal and bacterial counterparts (Prosser and Nicol, 2012) but found to be more representative of bacterial nitrifiers under NH₄⁺-N rich soil conditions (Hofferle *et al.*, 2010). Although these techniques provide a fairly good understanding about soil nitrifiers the anomalies common for culture based techniques should be considered when interpret results.

Intensive vegetable cultivation is practiced in several agro ecological regions in Sri Lanka in varying intensity and types of input management with different crop rotations leading to high diversity in vegetable grown systems. Therefore, we hypothesized that the abundance and the

activity of nitrifiers in soils will be different across different vegetable grown systems. This study was conducted to estimate the activity of nitrifiers and their abundance in soils from selected intensively vegetable cultivated regions using PNR and MPN techniques to generate preliminary information on dynamics of nitrifiers in these systems. The information will be useful in future research and planning activities to enhance N-fertilizer use efficiency and minimize NO_3^- leaching to groundwater. Further this study attempted to enumerate and isolate nitrifying bacteria for future research.

METHODOLOGY

Vegetable production under different input management systems in different agro-ecological regions were selected in order to introduce high diversity of systems to the study. Accordingly, four locations having contrasting climatic conditions and soil types, and also representing popular regions for vegetable production in the country, were selected. Introducing variability in input management, uncultivated lands under natural vegetation, vegetable cultivated fields managed conventionally using synthetic agrochemicals, and those managed with organic farming methods were included in the sampling scheme whenever possible. Details of the locations used for soil sampling is given in Figure 1 and Table 1.

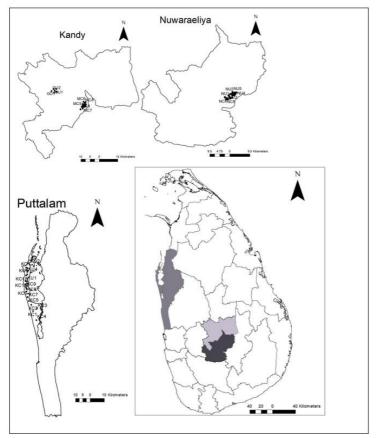


Figure 1. Map of Sri Lanka depicting sampling locations in three districts

Table 1. Details of the sampling locations and the number of samples collected from each location

	Location ¹						
	Nuwara Eliya	Marassana	Kalpitiya	Gannoruwa			
Agro ecological region	IU2	IM3c	DL3	WM2b			
Taxonomic class/classes of soil/s ²	Red yellow podsolic soils (Typic Paleudults)	Red yellow podsollic soils (Typic Paleudults)	Sandy regosols (Typic Quatzepsamments)	Reddish Brown latasolic (Typic Troporthents)			
Sampling localities	Bambarakele Seethaeliya, Mahagastota Katumana, Nuwara Eliiya	Marassana	Nawakkadu, Norochcholei, Kalpitiya, Narakkalliya	HORDI research fields			
Number: Conventional farm (CF) ³	6	18	18	3			
Organic farm (OF)	6	-	2	1			
Natural vegetation	5	6	5	2			
Types of crops in cultivated	CF: Carrot, beet, leeks	CF: Tomato, capsicum	CF: beet, onion, Chilies, pumpkin, batana, capsicum	CF: brinjal, tomato			
fields at the time of sampling	OF: Carrot cabbage broccoli	Luffa, cucumber, bitter guard	potato, pumpkin, Manioc, gherkin OF: Nivithi, beet root, long bean, okra, capsicum	OF: cabbage			

Sampling was performed for 4 months, from October, 2016 to March in 2017. Selected vegetable cultivated fields were having a crop at maturity stage at the time of sampling. From each field eight random soil samples were collected at 0-15 cm depth to make a composite sample representative of the field. Altogether 72 soil samples were collected for the study. Soil samples were immediately transferred to the laboratory at HORDI, Peradeniya, Sri Lanka. A subsample of the field moist soil was sieved through 4 mm sieve and refrigerated for subsequent microbial analyses. The remaining portion of each soil sample was air dried, sieved through 2 mm mesh sieve and stored for chemical and physical characterization.

Physico-chemical Analyses

Soil pH and electrical conductivity (EC) were measured in soil: water suspensions of 1:1 and 1:5, respectively. Soil organic carbon was analyzed using Walkley and Black method (Baruah and Barthakur, 1997) and active carbon content was analyzed according to permanganate

oxidiazable carbon (POXC) method described by Weil *et al.* (2003). Soil texture was determined by pipette method as explained in Dharmakeertihi *et al.* (2007).

Microbiological analyses

Soils were analyzed for total culturable bacteria, fungi, ammonia oxidizing bacteria and nitrite oxidizing bacteria using culture based techniques. Total culturable bacteria and fungi were enumerated on 3% triptic soy agar (TSA) and Rose Bengal agar, respectively, using serial dilutions in spread plate technique. Plates were incubated at 28±1 °C and bacterial colony counts were taken 2 days after inoculation and fungal colonies were counted 6 days after inoculation. Most probable number (MPN) method as described by Schmidt and Belser (1994) was used to estimate the abundance of ammonia oxidizers and nitrite oxidizers. Cultures were incubated in triplicates for 21 days before making observations, Potential Nitrification Rate (PNR) of soils was assessed following shaken slurry method (Heart et al., 1995). Optimum time points to collect subsamples for the determination of potential nitrification rates were identified after conducting a preliminary study. Accordingly subsamples of the slurry was drawn at 4 h and 20 h after initiating incubation at room temperature (26±2 °C) and nitrate was measured calorimetrically (Cataldo et al., 1975). Ammonia oxidizers were cultured in NH₄⁺ containing P buffer medium used in MPN method and plated on to TSA to obtain single colonies. Selected colonies were purified and screened for their ability to oxidize NH₄⁺. Isolates of ammonia oxidizers were cultured in NH₄⁺ containing P buffer medium and stored in refrigerator for future analyses.

Statistical Analysis

All the statistical analyses were performed using SPSS 16 statistical package. Data collected were used in preliminary descriptive analysis and Pearson correlation analysis for the measured parameters was performed at p<0.05. Principal component analysis (PCA) was performed to identify the variables among the measured parameters contributing mostly for the differences among soils. In here, correlation PCA was performed and the suitability of the data set for the structure detection was confirmed by performing Kaiser-Meyer-Olkin (KMO) Test and Bartlett's test of sphericity. Results indicated that sampling was adequate for factor analysis (KMO test value = 0.62) and there is adequate relationship between variables for structure detection (significance level at <0.001). Parallel analysis was performed to confirm the number of components to be retained from PCA and accordingly first three principal components (PC) were selected and single scores for each component for soil sample was calculated considering the variables contributing mostly to each PC to present data graphically.

RESULTS AND DISCUSSION

Soil samples collected from the four regions belonging to different climatic regions and having different soil types as indicated in Table 1 were diverse in terms of measured soil characteristics (Table 2). The pH of the studied soils ranged from very strongly acidic (3.8) to moderately alkaline (8.5). High variability in pH was observed even within the same management system in a region (e.g.CF in Nuwara Eliya) possibly due to multitude of factors including differences in topography and intensity of input management. Majority of the soil samples from Nuwara Eliya region were acidic while 92% of soils found in Kalpitiya were alkaline but none of the studied soils were saline (Table 2). Organic C percentages were high in 56% of the soil samples from Nuwara Eliya. In contrast, 60% of soils from Kalpitiya had

Table 2. Basic statistics of chemical properties of soil collected from Nuwara Eliya, Gannoruwa, Marassana and Kalpitiya

Location	GC	GO#	GU	KC	KO	KU	MC	MU	NC	NO	NU
Number	3	1	2	18	2	5	18	6	6	6	5
of											
samples											
pH											
Mean	6.6	6.7	6.3	7.6	7.8	7.7	6.6	6.9	5.6	6.5	4.9
St. Dev.	0.5	-	0.1	0.5	0.1	1.0	0.6	0.4	1.1	0.2	1.1
Minimum	6.2	-	6.3	6.4	7.7	6.0	5.2	6.1	4.5	6.3	3.8
Maximum	7.2	-	6.4	8.5	7.9	8.3	7.2	7.2	7.3	6.8	6.5
				F	EC (dS/	'm)					
Mean	0.19	0.06	0.08	0.12	0.13	0.25	0.25	0.18	0.37	0.2	0.12
a. <i>P</i>	0.02		0.004	0.04	0.02	0.14	0.22	0.02	0.2	0.10	0.00
St. Dev.	0.02	-	0.004	0.04	0.03	0.14	0.22	0.03	0.2	0.12	0.02
Minimum	0.17	-	0.08	0.06	0.10	0.09	0.04	0.10	0.12	0.13	0.09
Maximum	0.06	-	0.23	0.15	0.15	0.43	0.94	0.18	0.58	0.44	0.14
				Organ	ic Car	bon (%	5)				
Mean	1.9	1.2	4.0	0.7	0.9	1.5	1.5	1.5	2.7	3.2	3.1
St. Dev.	0.5	-	0.5	0.8	0.1	1.3	1.1	1.1	0.6	0.8	0.5
Minimum	1.6	-	3.6	0.3	0.9	0.4	1.0	0.5	1.6	2.3	2.3
Maximum	2.4	-	4.3	3.8	1.0	3.4	3.2	3.0	3.4	4.5	3.7
Active carbon (mg/kg)											
Mean	248	261	207	257	395	457	307	419	290	371	563
St. Dev.	233	-	46	153	49	263	111	126	155	213	138
Minimum	34	-	174	2	360	135	120	284	103	136	419
Maximum	496	-	240	306	429	815	453	630	553	708	749

Nuwara Eliya, Gannoruwa, Marassana and Kalpitiya are denoted by N, G, M and K, respectively as first letter in data labels representing uncultivated lands and fields cultivated with vegetables using conventional and organic practices (as denoted by U, C and O, respectively as second letter in data labels). There was only one field sampled from Gannoruwa representing organic vegetable production. Therefore, standard deviation, minimum and maximum are not calculated

very low organic C content. In the present study, there was no clear effect of implementing organic management practices on soil organic C contents. This may be partly due to the less number of samples representing organically cultivated fields compared to conventionally managed fields. Active carbon quantifies labile carbon in soil (Culman *et al.*, 2013). The amount of labile carbon influences the availability of readily utilizable carbon to microorganisms and hence, their activity (Hurisso *et al.*, 2016). It is considered as a sensitive indicator of soil quality (Culman *et al.*, 2013). The active carbon contents varied widely in analyzed soil samples (Table 2) and the values are within the range reported previously, which is from 24 to 1469 mg kg⁻¹ (Culman *et al.*, 2013; Hurisso *et al.*, 2016). Soil texture was highly variable among the tested samples with sand contents ranging from 6 to 97%. In overall, the most prominent textural classes of the studied soils were Sandy clay loam (34%) and Sandy loam (30%). The highest diversity in soil texture was found in samples collected from Marassana with Sandy clay loam, Sandy loam, Loam, Loamy sand and Sandy clay as the textural classes. Soil samples collected from Gannoruwa had Sandy clay loam texture. The texture of soil samples collected from Nuwara Eliya were mostly loamy type and belonged to

textural classes of Sandy clay loam, Sandy loam, Silt loam, Loam and Clay loam. On the other hand, the soils collected from Kalpitiya region were mostly sandy type and belonged to textural classes of sandy clay loam, sandy loam, loamy sand and sand.

The potential activity of nitrifiers ranged from 0.1 to 16.0 NO₃⁻-N mg/kg of dry soil/ h (Figure 2a). Mostly conventionally managed fields showed higher PNR rate compared to other cropping histories in respective regions. This may be due to the higher ammonium levels applied by farmers. When analyzing total ammonia oxidizing bacteria and nitrite oxidizing bacteria referred to as nitrifying bacteria, above results cannot be explained subjected to population size of them (Figure 2b and 2c). The efficiency of nitrifying communities under optimum conditions when expressed as potential nitrification rate per unit of ammonia oxidizers, who are conducting the rate limiting first step of nitrification, varied from 1 to 57,341 pg NO₃⁻-N/ ammonia oxidizer cell/ h suggesting differences in species composition in these communities

Table 3. The abundance of total culturable bacteria and fungi in the soil collected from Nuwara Eliya, Gannoruwa, Marassana and Kalpitiya

Location	NO	NC	NU	GO	GC	GU	MC	MU	КО	KC	KU
Bacteria (log10 CFU g ⁻)											
Mean	8.1	7.9	7.8	-	7.7	8.35	8.2	8.1	8.8	8.6	8.9
St. Dev.	0.3	0.2	0.2	-	0.2	0.01	0.2	0.2	0.1	0.2	0.5
Minimum	7.69	7.7	7.6	-	7.6	8.34	7.5	7.8	8.7	8.1	8.6
Maximum	8.5	8.2	8.1	-	7.9	8.36	8.5	8.4	8.9	9.0	9.9
			Fun	gi (log	10 CF	U g ')					
Mean	4.1	3.7	3.6	-	3.2	3.3	3.3	4.9	4.5	4.5	4.4
St. Dev.	0.8	0.5	0.2	-	0.2	0.0	0.3	0.10	0.1	0.4	0.1
Minimum	3.3	3.1	3.4	-	3.0	3.3	2.7	4.8	4.4	3.6	4.2
Maximum	5.2	4.4	3.9	-	3.0	3.3	3.9	5.1	4.5	4.9	4.5

Nuwara Eliya, Gannoruwa, Marassana and Kalpitiya are as denoted by N, G, M and K, respectively as first letter in data labels representing uncultivated lands and fields cultivated with vegetables using conventional and organic practices (as denoted by U, C and O, respectively as second letter in data labels)

Loamy type textures usually facilitate good air and water balance while soils with sandy type textures are highly aerated and have less water retention. When water retention in soil is poor it affects substrate diffusion to microbial cells and cause adverse physiological effects due to cell dehydration; thus, affecting nitrification (Stark and Firesotne 1995). Further, oxidation of organic matter is high in well aerated soils than poorly aerated soils when moisture level is not limiting microbial activity and therefore, organic matter retention is less in sandy type texture compared to loamy or clayey type soil textures. Significant negative correlations between sand content of soils and the organic carbon and active carbon contents were seen in the present study (Table 4). Soil pH correlated with a number of parameters measured (Table 4). Therefore, the nitrifying communities of the studied soils may have evolved under unique sets of environmental constraints as facilitated by interacting effects of soil properties like texture, pH and organic carbon contents with agro-climatic conditions (O'Sullivan *et al.*, 2013).

Population size of nitrite oxidizers were higher compared to ammonia oxidizers (Figure 2b. and 2c.). Sergei Winogradsky first isolated and showed in 1890 the organisms responsible for two steps of nitrification. 1st step is carried out by mainly *Nitrosomonas, Nitrosospira, Nitrosolobus,* and *Nitrosococcus*. Second step is carried out by mainly *Nitrobactor,* followed by *Nitrospira* and *Nitrosococcus* (Sullivan *et al.,* 2012). However, recent studies indicated that some members of genera *Nitrospira* could perform complete oxidation of NH₄⁺ to NO₃⁻ via comammox (Kessel *et al.,* 2015).

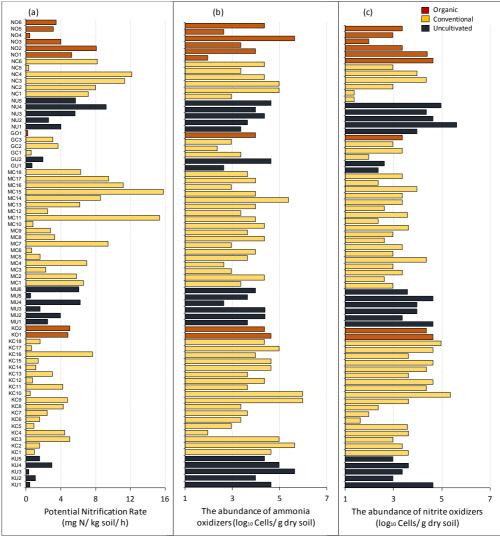


Figure 2. The activity and abundance of nitrifiers as indicated by (a) potential nitrification rate (mg of nitrate N/kg of dry soil /h) (b) abundance of ammonia oxidizers and (c) abundance of nitrite oxidizers of the soils collected from Nuwara Eliya, Gannoruwa, Marassana and Kalpitiya (as denoted by N, G, M and K, respectively as first letter in data labels) representing uncultivated lands and fields cultivated with vegetables using conventional and organic practices (as denoted by U, C and O, respectively as second letter in data labels)

Nitrospira could be found in high abundance in soil but their activity is very slow compared to *Nitrosomonas* like species (van Kessel *et al.*, 2015). Further, AOB like *Nitrosomonas* spp.

frequently found in agricultural soils are metabolically classified as obligate chemolithoautotrophs that utilize NH₄⁺ as an energy source and CO₂ as the C source. Increase in activity of this organism in the presence of organic substrates has been reported (Hommes *et al.*, 2003). Therefore, the activity of the same group of organism could be different with respect to soil environmental characters such as composition and availability of labile carbon.

Techniques used in the present study to analyze PNR and the abundance of nitrite oxidizers and ammonia oxidizers do not discriminate between bacterial and archaeal counterparts (Prosser and Nicol, 2012). However significant differences in terms of the contribution of these two groups of organisms for nitrification have been suggested based on molecular based approaches in a number of studies (Leininger et al., 2006; Bernhard et al., 2010; Hofferle et al., 2010; Posser and Nicol 2012; Hu and He, 2014). Bacterial nitrifiers are reported to dominate nitrification under slightly acidic to neutral pH soil conditions and also under environments receiving high NH₄⁺ inputs (Fan et al., 2011; Tylor et al., 2011; Verhamme et al., 2011). According to Tylor et al. (2011) fertilizer-N application and manure addition increased AOB abundance in soils compared to unfertilized soils while AOA abundance remained unchanged. However, the community composition of both AOB and AOA groups changed with fertilizer treatments. They further reported that change in community composition of AOB with crop growth stage was more prominent than that of AOA. Therefore, there should be high variability in the dominance of AOB and AOA in the studied soils that represent uncultivated and cultivated systems, which are different with respect to input management and vegetation.

Higher number of archaeal nitrifiers have been reported under extreme soil conditions for microbial growth such as high or low pH, salinity, etc. (Bernhard et al., 2010; Hofferle et al., 2010; Posser and Nicol 2012; Hu and He, 2014). Change in abundance of these two groups depending on temperature and soil moisture regimes and altitude has also been reported (O'Sullivan et al., 2013; Tylor et al., 2011; Zhang et al., 2009). The altitude of the sampling sites used in the present study varied from 8 m (e.g. Kalpitiya) to above 1860 m (e.g. Nuwara Eliya) from mean sea level and along with the elevation gradient the climatic conditions the soils are experiencing changes as indicated by the different agro-ecological regions the locations belong to (Table 1). Further high diversity in soil texture is suggestive of high variability in temperature and moisture conductivity among many other processes affecting microbial growth. Therefore, it can be expected that even though two soils have relatively similar abundance of ammonia oxidizers the activity can be very different because composition of community affects the potential activity. This is evident in the present study since the abundance of ammonia oxidizers in soils collected from Kalpitiya and Marassana are less variable compared to the variability exist in the potential activity of nitrifiers from the same soils (Figure 1a and 1b). Potential nitrification rate is often reported to show correlation with AOB abundance than AOA (Bernhard et al., 2010). Lack of correlation between PNR and AOA and AOB abundance was observed by O'Sullivan et al. (2013). In the present study, significant correlation was observed between PNR and EC but there was no correlation between PNR and the abundance of either ammonia oxidizer or nitrite oxidizer groups (Table 4). There may be differences in terms of the contribution of AOA and AOB for the nitrification in soils used in present study.

Although there is high variability in the estimated characters of the nitrifying populations these characters were not strong at defining the variability of the studied soils as indicated by PCA (Table 5 and 6). The variability among the studied soils was captured by PC1, PC2 and PC3 at 30, 16 and 14%, respectively. Structure of data along PC1 is mostly due to soil pH and

organic C and abundance of bacteria and fungi. Potential nitrification rate and EC scored high in PC2. O'Sullivan *et al.* (2013) reported that PNR was more sensitive to the soil fertility status than the population sizes of AOA and AOB. In the present study PNR contributed for the variability of studied soils than ammonia oxidizer or nitrite oxidizer abundance as depicted by PCA.

Despite the differences in soil characteristics and agro-climatic conditions, the studied vegetable production systems have one character in common, which is intensive production with high input use (organic and/or synthetic fertilizers). Therefore, the soils receive high dose of nitrogenous fertilizers in each cropping cycle. The intensity of cultivation considering the number of crops per year vary as NuwaraEliya>Kalpitiya>Marassana=Gannoruwa. These regions are environmentally sensitive areas as Nuwara Eliya, Gannoruwa and Marassana are in tributaries of major rivers (mainly river Mahaweli) and Kalpitiya is located on a perched water table. Therefore, contamination of groundwater and surface water bodies with nitrate causing on-site and off-site effects is highly probable in these regions.

Table 4. Correlations between measured variables

	pН	EC	OC	AC	PNR	TB	TF	AO	NO
EC	-0.137								
	NS								
OC	-0.490	0.126							
	***	NS							
AC	-0.175	-0.039	0.183						
	NS	NS	NS						
PNR	-0.225	0.547	0.193	0.109					
	*	***	NS	NS					
TB	0.500	-0.060	-0.339	0.034	-0.237				
	***	NS	**	NS	*				
TF	0.296	-0.214	-0.297	0.044	-0.389	0.361			
	*	*	*	NS	**	**			
AO	0.200	-0.075	-0.175	0.163	-0.084	0.282	0.329		
	NS	NS	NS	NS	NS	*	**		
NO	0.157	-0.263	-0.096	0.203	-0.116	-0.021	0.021	0.146	
	NS	*	NS	NS	NS	NS	NS	NS	
sand%	0.49	0.001	-0.51	-0.496	0.014	0.004	0.18	-0.11	-0.16
	***	NS	***	***	NS	NS	NS	NS	NS

OC-organic carbon, AC-active carbon, PNR-potential nitrification rate, TB-total bacteria, TF-total fungi, AO-ammonia oxidizers, NO-nitrite oxidizers

Significance at probability levels: p<0.05, *: p<0.01, **: p<0.001, ***

Table 5. Total Variance explained by the first three principal components (PCs)

Component	Initial Eigenvalues								
	Total	% of Variance	Cumulative %						
1	2.758	30.645	30.645						
2	1.435	15.940	46.585						
3	1.290	14.331	60.916						

NS - Not significant:

Table 6. Scores of each variable in Rotated Component Matrix

Variable	Component						
variable	1	2	3				
Bacteria abundance	0.797	0.013	0.126				
pН	0.751	-0.119	-0.109				
Organic C	-0.683	0.106	0.202				
Fungi abundance	0.668	-0.243	0.120				
EC	-0.042	0.877	-0.072				
Potential nitrification rate	-0.280	0.796	0.101				
Active C	-0.176	0.062	0.829				
Ammonia oxidizer abundance	0.483	0.043	0.566				
Nitrite oxidizer abundance	-0.005	-0.423	0.497				

There are enough evidences to indicate nitrate pollution in surface and ground water due to fertilizer usage in vegetable production is already an issue (Henegama *et al.*, 2013; Kuruppuarachchi, 2010; Jayasingha *et al.*, 2011; Liyanage *et al.*, 2000). Therefore, information on potential activity of nitrifiers and the diversity of nitrifying communities are important to design suitable techniques to manage N fertilizers and minimize N losses to the environment.

From among the parameters estimated in relation to nitrifiers, PNR was the only parameter that contributed to the structured variability among soils (Table 5 and 6). Based on the PNR values observed in the present study about 48% of soils had PNR values from 3 to 10 mg of N/kg soil/h and 7% of soils had values greater than 10 mg of N/kg soil/h. Olsson and Falkengren-Grerup (1999) reported PNR in soils from Oak forests in Southern Sweden ranging from 0 to 24 nmol NO₃-/g/h which is equivalent to 0 to 0.34 mg of N/kg/h. They further reported that PNR positively correlated with pH. Tylor et al. (2011) reported PNR rates as high as 2.4 µmol N/g/day (equivalent to 1.46 mg of N/kg soil/h) when soil is incubated with 1 mM NH₄⁺ supplement. Based on a study conducted on an Alfisol soil in Northeast China, Fan et al. (2011) indicated that PNR tend to be suppressed by long term mineral fertilizer application but enhanced by application of manures such as horse manure. The reported PNR rates by Fan et al. (2011) ranged from 0.5 to 2.6 mg NO_3 -/kg soil/h (equivalent to 0.11 to 0.59 mg of N/kg soil/h). Therefore, PNR in most of the soils in this study are higher than those reported in literature (Figure 2a). Considering the high PNR values observed in the present study in agricultural soils there is a high potential of losing N once nitrogenous fertilizer are added to cultivated fields, contaminating groundwater and surface water with NO₃-, and increasing emissions of N₂O, a greenhouse gas with 296 times more global warming potential than CO₂ (Dalal et al., 2003). The N₂O is emitted during nitrification to some extent and largely through denitrification of NO₃ produced from nitrification (Dalal et al., 2003). Hence timely actions are needed to be taken to moderate nitrification rate as vegetable grown regions receive high quantities of nitrogen fertilizers. Nine ammonia oxidizer isolates were made from the studied soils to facilitate future research activities on developing suitable techniques to suppress nitrification in vegetable cropping systems.

In the present study, the abundance of bacteria, fungi, ammonia oxidizers and nitrite oxidizers are presented in per gram of soil basis, which is a standard expression used. However, differences in patterns can be expected if the same has been presented in per unit soil volume because the soil texture and organic matter content vary widely across soils and structural properties, especially bulk density, also vary across fields (Mehlich, 1972). Since soil bulk density measurements were not made the comparison in per volume basis cannot be performed.

It is suggested that in future research of this nature it will be important to consider expressing results in per volume of soil.

CONCLUSIONS

The soil characteristics, potential activity and abundance of nitrifiers' populations varied widely across the studied soils. The results indicate very high diversity in nitrifying communities inhabiting intensively vegetable cultivated soils. PNR contributed more to structuring the variability of soils than abundance of ammonia oxidizers and nitrite oxidizers, thus, it is a better indicator of dynamics of nitrifiers than population size of the group. It is revealed that potential nitrification rates of 55% of soils from the studied vegetable grown regions were higher than the values reported in literature. Hence, moderation of nitrification rate should be crucial, especially in vegetable grown soils, to improve agronomic fertilizer use efficiency and reduce groundwater contamination with nitrate and emissions of N₂O as these fields receive high amounts of nitrogen fertilizers with each cropping cycle. The variability in soils with respect to functional diversity of nitrifying communities may have to be considered when developing mechanisms for the moderation of nitrification.

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